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**particle properties
related to black
carbon in Mexico City**

D. Baumgardner et al.

On the diurnal variability of particle properties related to black carbon in Mexico City

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Title Page

Abstract

Introduction

Conclusions

References

Tables

Figures

◀

▶

◀

▶

Back

Close

Full Screen / Esc

Printer-friendly Version

Interactive Discussion

Abstract

The black carbon mass (BCM) of individual, internally mixed aerosol particles was measured with the Single Particle Soot Photometer (SP2) in April of 2003 and 2005. The average BCM, single particle BC mass fraction and BCM equivalent diameter were evaluated with respect to concentrations of carbon monoxide (CO), particle bound polycyclic aromatic hydrocarbons (PPAH) and condensation nuclei (CN). The BCM and CO have matching diurnal trends that are linked to traffic patterns and boundary layer growth. The PPAH reaches a maximum at the same hour as CO and BCM but returns rapidly back to nighttime values within three hours of the peak. The number of particles containing BCM ranges between 10% to 40% of all particles between 200 nm and 700 nm and the BCM is between 4% and 12% of the total mass in this size range. The average BC equivalent mass diameter varies between 300 and 400 nm and reaches its daily minimum value when BCM is a maximum. The BC particles have the thinnest coating of non-light absorbing material during periods of maximum BCM. The scattering and absorption coefficients, B_{scat} and B_{abs} , derived from the SP2 measurements were compared with direct measurements from a nephelometer and soot photometer. The measured and derived B_{abs} are in close agreement whereas the B_{scat} comparisons show larger discrepancies in absolute value and daily trends. Even though approximately 40% of the BCM is in particles with diameters smaller than 200 nm, the extinction coefficient is dominated by the BCM in particles larger than this size. The BCM contributes up to 20% of the total extinction in this size range. BCM is emitted at a rate of 1200 metric tons per year in Mexico City, based upon the SP2 measurements and correlations between BCM and CO.

1 Background

The environmental impact of atmospheric aerosols is multifaceted as a result of the wide range in their concentration and size, complexity of chemical composition and

ACPD

7, 1623–1653, 2007

particle properties related to black carbon in Mexico City

D. Baumgardner et al.

Title Page

Abstract

Introduction

Conclusions

References

Tables

Figures

◀

▶

◀

▶

Back

Close

Full Screen / Esc

Printer-friendly Version

Interactive Discussion

EGU

optical properties that contribute to their role in global climate change (e.g., Charlson et al.1992), photochemistry (e.g., Dickerson et al., 1997; Jacobson, 1998; Raga and Raga, 2000) and cloud processes (e.g., Hobbs, 1991). Black carbon aerosol (BCA)¹ is an anthropogenic component of atmospheric particles that has been recognized for its environmental impact for more than 100 years (Novakov, 1982). These ubiquitous particles, found throughout the world's atmosphere, are produced primarily by incomplete combustion of fossil fuel and from burning of biomass (e.g., Ghan and Penner,1992) and may have much longer lifetimes than other types of aerosol particles.

The climatic importance of BCA is a result of the high efficiency with which they scatter and absorb light over a wide range of solar and infrared wavelengths. This is highlighted by comparing the efficiency with which radiant energy is removed by BCA and ammonium sulfate, another type of atmospheric particle found in abundance in urban and rural areas. Figure 1 illustrates the difference in extinction efficiency between these two types of particles at 550 nm, the wavelength of maximum solar radiance. The extinction was calculated incorporating Mie theory (Bohren and Huffman, 1983) and assuming spherical particles with the refractive indices of ammonium sulfate (1.43–0.0i) and BCA (1.750–0.440i) at this wavelength (Twitty and Weinman, 1971; Volz, 1972).

For diameters smaller than 30 nm, BCA have an extinction efficiency that is larger

¹ As noted by Schwarz et al. (2006) the term “black carbon” has different meanings depending on the context of the discussion and the method of detection. It is common to associate the meaning of measurements of carbonaceous aerosol components with the specific instrument being used (Pöschl, 2002). The measurements discussed in the current paper are from an instrument that detects only the refractory and strongly light-absorbing component of combustion generated aerosols. This component is referred to as either elemental carbon (EC) or BC depending on whether particles are being classified by their thermochemical or optical properties, respectively. Thus EC and BC are often used interchangeably. In the present study we will follow the convention used by Schwarz et al. (2006) and label the ambient measurements as black carbon aerosol (BCA), recognizing that BCA represents only a component of the carbonaceous material often called soot.

**particle properties
related to black
carbon in Mexico City**

D. Baumgardner et al.

Title Page

Abstract

Introduction

Conclusions

References

Tables

Figures

◀

▶

◀

▶

Back

Close

Full Screen / Esc

Printer-friendly Version

Interactive Discussion

than sulfate by factors of up to several orders of magnitude and is comparable at larger diameters. Although the number concentration of BCA in the free troposphere is generally less than that of sulfate aerosols their impact on radiative fluxes can be of the same order of magnitude. Furthermore, as shown by recent modeling and laboratory studies (Bond et al., 2006; Mikhailov et al., 2006) coatings of non-light absorbing material on the surface of BCA will enhance the extinction efficiency by factors of 1.5 to 2.5 as a result of multiple reflections within the coating on the BC.

The optical properties of BCA must be taken into account when interpreting measurements by remote sensors, i.e. lidars, sun photometers and satellites. The algorithms that are applied for extracting information from these measurements incorporate assumptions about the aerosol properties, e.g. the shape of the size distribution and the composition and optical characteristics. In some cases, the objective is to derive particle properties whereas in other applications aerosols are a source of interference to be removed. Many years of airborne and ground based measurements have gone into the parameterization of aerosol properties to use in these algorithms; however, until recently, very few measurements had been made of BCA characteristics.

The study described herein is a particle by particle evaluation of BCA properties in Mexico City, their diurnal trends and links to other related anthropogenic gases and aerosols. The remainder of the paper describes the instrumentation, evaluates the physical and optical characteristics of BCA and discusses the relationships between primary emissions and meteorological processes with respect to the aerosol properties.

**particle properties
related to black
carbon in Mexico City**

D. Baumgardner et al.

Title Page

Abstract

Introduction

Conclusions

References

Tables

Figures

◀

▶

◀

▶

Back

Close

Full Screen / Esc

Printer-friendly Version

Interactive Discussion

2 Measurement and Analysis Methodology

The measurements were made with a Single Particle Soot Photometer² (SP2), a nephelometer and Particle Soot Aerosol Photometer³ (PSAP), a Photoelectric Aerosol Sensor⁴ (PAS-2000) and with a condensation nucleus (CN) counter (TSI Model 3010⁵).

- 5 These instruments provide measurements of the BCA physical, chemical and optical properties, as described below in Sect. 3 and 4.

3 BC physical properties

10 The SP2 employs a patented technique (Stephens et al., 2003) that combines the principles of light scattering, absorption and emission to derive the diameter, mass and incandescence temperature of individual aerosol particles in the diameter range (for the current study) from 0.20 to 0.70 nm (Baumgardner et al., 2004; Schwarz et al., 2006; Moteki and Kondo, 2006). Particles enter the sample cavity of the SP2 (the typical sample flow rate for the instrument is $2\text{ cm}^3\text{ s}^{-1}$) and while passing through the beam of a diode-pumped Nd:YAG laser ($1.064\text{ }\mu\text{m}$ wavelength) they scatter and
15 absorb light. Two cones of scattered light, 30° – 60° and 120° – 150° , are collected by the optics and focused on a photodetector that produces a voltage signal proportional to the scattering intensity. Particle diameter is derived from the scattered light using classical Mie theory.

20 Particles that contain material that absorbs light at the wavelength of the laser are heated and reach a temperature at which they incandesce and emit light at a wavelength that is a function of the temperature. Two detectors, each with filters for passing different wavelengths, measure the emitted light. The temperature is derived from the

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Title Page

Abstract

Introduction

Conclusions

References

Tables

Figures

◀

▶

◀

▶

Back

Close

Full Screen / Esc

Printer-friendly Version

Interactive Discussion

ratio of signals from these detectors (Baumgardner et al., 2004; Schwarz et al., 2006; Moteki and Kondo, 2006). The intensity of the emitted light is proportional to the mass of the light absorbing material and the temperature of incandescence identifies the composition.

5 Calibration of the SP2 is carried out with commercially available spherical particles. Monodispersed, polystyrene latex spheres are used to calibrate the scattering signals and glassy carbon spheres of known density were size-selected with an electrostatic classifier for calibration of the incandescence signals. Incandescing particles are sized in the range of 3–300 fg/particle (200–700 nm mass equivalent diameter at 1.42 g cm⁻³ density). The optical diameter between 200–700 nm is derived from the scattering signal. The scattering and incandescence signals are recorded particle-by-particle such that the instrument is sensitive to very small concentrations of BCA.

10 The black carbon mass (BCM) fraction, i.e. the ratio of BCM to total mass of an individual, internally mixed particle, is estimated by comparing the mass equivalent diameter of the BC particle with the optical diameter:

$$\text{BCM}_{\text{fraction}} = (D_{\text{mass_eqv}}/D_{\text{optical}})^3 \quad (1)$$

and

$$D_{\text{mass_eqv}} = (6\text{BCM}/\pi\rho)^{1/3} \quad (2)$$

20 A value 2.24 g cm⁻³ is assumed for the BC density (Fuller et al., 1999). An advantage of the SP2's optical design is that the collection of scattered light over a wide solid angle decreases the sensitivity of the derived particle diameter to variations in the refractive index. As shown in Fig. 2, using the ammonium sulfate and BC particles as an example, the scattering cross sections are similar even though the refractive indices are quite different. The refractive index selected for the current study assumes an average mixture of 10% BC and 90% organo-sulfate by mass. The average refractive index for this mixture is 1.48–.05i. The choice of this refractive index limits the uncertainty in derived diameter to a maximum of ±0.03 μm as shown by the box drawn in Fig. 2. It

**particle properties
related to black
carbon in Mexico City**

D. Baumgardner et al.

Title Page

Abstract

Introduction

Conclusions

References

Tables

Figures

◀

▶

◀

▶

Back

Close

Full Screen / Esc

Printer-friendly Version

Interactive Discussion

is important to note that the measured BCM is insensitive to the depth of the non-light absorbing coating (Slowik et al., 2006; Moteki and Kondo, 2006).

The number concentration of Mexico City's aerosol are dominated by particles smaller than the SP2's threshold of 200 nm as previous measurements with a differential mobility analyzer (DMA) have shown (Baumgardner et al., 2000); however, more than 60% of the aerosol volume is in particles larger than this diameter. Furthermore, evaluation of BCM on aluminum substrates exposed in an eight stage, multi-orifice, uniform deposit impactor (MOUDI), using evolved gas analysis (Baumgardner et al., 2002; Raga et al., 2003), indicate that almost 80% of the BCM is in BCA larger than 180 nm. Following the analysis approach taken by Schwarz et al. (2006), we estimate the fraction of BCM that is below the size detection limit of the SP2 by fitting a lognormal curve to the mass size distribution of BCA derived from accumulated SP2 measurements. The volume size distributions derived previously from DMA measurements in Mexico City (Baumgardner et al., 2000) have a shape that is lognormally distributed. Lognormal distributions of BCA have also been observed in the free troposphere by Clarke et al. (2004) and in other urban environments by (Kondo et al., 2006). Figure 3 illustrates a typical BCM size distribution and the monomodal lognormal function that is fit to it. The lower horizontal axis represents the mass-equivalent diameter of BCA as defined in (2). Based on the comparison of the BCM measured with the SP2 and the derived mass found by integrating the lognormal functions, approximately 60% of the BCM is in equivalent diameters larger than $200\text{ }\mu\text{m}$ (shaded area under the solid curve). In the remainder of this paper, the bulk BCM values that are reported and used in comparisons with the other atmospheric parameters are derived from the lognormal fit to hourly averages of the BCM size distributions.

A detailed description of the theory of operation, uncertainties in determining the size and mass of BCA and measurement limitations are documented elsewhere (Schwarz et al., 2006; Moteki and Kondo, 2006). For the interpretation of the measurements presented in the present study, the estimated uncertainties in BCM, D_{optical} , $D_{\text{mass_eq}}$ and $\text{BCM}_{\text{fraction}}$ are $\pm 20\%$, $\pm 10\%$, $\pm 30\%$ and $\pm 36\%$, respectively.

**particle properties
related to black
carbon in Mexico City**

D. Baumgardner et al.

Title Page

Abstract

Introduction

Conclusions

References

Tables

Figures

◀

▶

◀

▶

Back

Close

Full Screen / Esc

Printer-friendly Version

Interactive Discussion

4 BCA Optical Properties

The scattering, absorption and extinction coefficients, B_{scat} , B_{abs} and B_{ext} , at an incident wavelength of $\lambda=550$ nm, are derived for individual particles measured by the SP2 using optical efficiencies calculated from Mie theory (Bohren and Huffman, 1983) and assuming that the particles are spherical. A look-up table was generated that contains refractive indices for particles in the size range of the SP2 and with 21 mixtures of BCA with a refractive index of $1.75-0.44i$ and organo-sulfate with refractive index of $1.48-0.00i$. The fraction of BCA by volume in a particle is varied between 0 and 1 in 0.05 steps and the refractive index of the mixture is computed using the volume mixing rule (Bohren and Huffman, 1983). For example a particle containing 10% BCA by volume will have a refractive index of $(0.9*1.48+0.1*1.75)-0.1*0.44 i$.

The scattering, absorption and extinction coefficients and the single scattering albedo are computed by determining the BCM_{fraction} of each particle and finding the optical parameters associated with this fraction in the lookup table.

An independent measurement of B_{abs} was made with the PSAP, an instrument that uses the integrating plate technique (Bond et al., 1999). Particles are collected on a filter and the transmission of green light through this filter is measured in real time. By definition, the value of B_{abs} is

$$B_{\text{abs}} = (A/V)\ln(I/I_0) \quad (3)$$

where I and I_0 are the transmitted and incident light intensities, A is the area of the filter covered by aerosol particles and V is the volume of air that passed through the filter to deposit a given layer of particles. The source of incident radiation in the PSAP is a light emitting diode ($\lambda=550$ nm) and particles are collected on a 10 mm quartz filter. The light is transmitted through a second filter, unexposed to ambient air and is the reference for the incident radiation, I_0 . Detailed laboratory studies of these instruments (Bond et al., 1999) have arrived at an empirical correction to Eq. (3) that accounts for the scattering of the aerosol on the collection filter.

$$B_{\text{abs}} = (B_{\text{abs_measured}} - 0.02B_{\text{scat_measured}})/1.2 \quad (4)$$

1630

ACPD

7, 1623–1653, 2007

particle properties related to black carbon in Mexico City

D. Baumgardner et al.

Title Page

Abstract

Introduction

Conclusions

References

Tables

Figures

◀

▶

◀

▶

Back

Close

Full Screen / Esc

Printer-friendly Version

Interactive Discussion

EGU

$B_{\text{abs_measured}}$ is the absorption coefficient derived from (3) and $B_{\text{scat_measured}}$ is the scattering coefficient measured with a nephelometer. This equation corrects for the effects of light scattering from particles on the filter, an interference that decreases the transmission through the filter, and for matrix effects of the quartz fibers that also enhance the measured attenuation and lead to overestimates of B_{abs} . Additional corrections are made for variations in the diameter of the aerosol deposit and changes in flow velocity related to the differences in the altitude of Mexico City and Seattle, Wa, where the instrument flow meter was calibrated.

5 Particle chemistry

Particle-bound polycyclic aromatic hydrocarbon (PPAH) was measured with the PAS 2000 that uses 207 nm light from an Excimer lamp with energies of 6.7 eV. Compounds on the surface of particles that have ionization potentials below the energy of the incident photons will release an electron that is subsequently removed with an ion trap. The positively charged particle is collected on a filter that is connected to an electrical ground and the electrical potential between the ion trap and positively charged filter creates an electrical current, proportional to the number of ions that reach the filter per unit time and is measured with an electrometer. The ionization potentials of PPAH on the surfaces of particles are generally below the energy of the incident photons; however, the PAS is also sensitive to BCA that has a work function of 4.4 eV. Some other organic compounds can also be ionized (Marr et al., 2006) and contribute to the signal from the PAS; hence, the signal from the PAS should be considered a relative, rather than absolute indicator of the concentration of mixtures of PPAH, BC and some trace amounts of other organics. The manufacturer cautions that the PAS must be carefully calibrated with the aerosol mixture being studied if the results are to be interpreted quantitatively. This type of calibration is difficult, if not impossible, for an urban environment where the aerosol mixture will change significantly depending on the relative fractions of BCA and organics. Given the uncertainties in absolute value, the results

Title Page

Abstract

Introduction

Conclusions

References

Tables

Figures

◀

▶

◀

▶

Back

Close

Full Screen / Esc

Printer-friendly Version

Interactive Discussion

from the PAS will be reported as a mass concentration of PPAH*, where the asterisk indicates that this quantity is only a relative indicator of PPAH.

Carbon monoxide was not part of the instrument package used in the current study; however, it is a useful tracer for primary emissions and can also indicate changes in the height of the boundary layer (Baumgardner et al., 2003). The CO used in the following analysis is from measurements made with the Mexico City pollution monitoring network, Red Automatica de Monitoreo Ambiental (RAMA). Two RAMA stations are within five kilometers of the measurement site discussed below. The hourly CO values from these stations were averaged to obtain the data reported here.

6 Results

The measurement site is on the roof of the building of the Centro de Ciencias de la Atmósfera (CCA) inside the campus of the Universidad Nacional Autónoma de México (UNAM). The university is located in the southwest quadrant of the Mexico City basin and the CCA building (19.6° N, 99.10° W) is 2288 m above sea level. The university campus is in a residential area with no significant industrial or commercial activities; however, nearby there is a steady flow of auto and bus traffic related to the circulation of university staff and students. Approximately 400 m to the southeast is the subway terminal where more than 400 buses collect passengers daily (Peralta et al., 2006).

The particle instrumentation was connected to a common manifold that draws its air from a fifteen meter chimney ventilated at a rate of approximately 90 l min⁻¹. All the instruments except the SP2 were operated 24 h a day. The data reported here are from 27 April to 1 May 2003 and 2–8 April 2005. The same SP2, CN counter, PSAP and nephelometer were used in both years. The SP2 was only operated from 07:00 to 18:00 in 2003 but made measurements 24 h a day in 2005. The PPAH* results reported here are only from the 2005 campaign. Unless noted otherwise, the data discussed below represent a compilation from both years, with the exception of the PPAH*.

The CN concentration, B_{scat}, and B_{abs} were output to the data system at 1 Hz and

**particle properties
related to black
carbon in Mexico City**

D. Baumgardner et al.

Title Page

Abstract

Introduction

Conclusions

References

Tables

Figures

◀

▶

◀

▶

Back

Close

Full Screen / Esc

Printer-friendly Version

Interactive Discussion

the PPAH* is an average over six seconds. The SP2 records information on every particle; however, as a result of the very high data rates and large quantity of data, this instrument was set to record data during one minute periods out of every ten minutes. All of the parameters were averaged over ten minute intervals. The RAMA CO data were only available as hourly averages.

Examination of the time series for CN, B_{abs} , BCM and CO indicates a reproducible evolution with the daily maxima occurring at approximately the same time each day. Figures 4a-f display the diurnal patterns of CO, CN, PPAH*, B_{abs} and B_{scat} averaged over the entire sampling period (vertical bars indicate the standard deviation about the mean). Note that the values of CO, CN, PPAH* and B_{abs} all increase rapidly at approximately 6am local time from very low and fairly constant nighttime values. Larger day-to-day variability, represented by the standard deviations, is evident in all parameters until sunset (~ 1900 LST) except for PPAH* which decreases significantly both in magnitude and variability after 1000. All these variables, except for B_{scat} , are indicators of primary emissions. Note that B_{scat} follows a different diurnal pattern whereby it reaches a maximum later in the morning compared to the other variables. This particle property is more sensitive to the evolution of particle sizes and is loosely linked to primary emissions.

Figures 5a-f show the daily averages of the BCA number and mass concentration, D_{mass_eqv} , $BCM_{fraction}$ and the ratios of the bulk BCA concentration and mass to the total concentration and mass. As in Fig. 4, the time series in each panel represent averages over the 14 days. All of the variables show similar trends that are dominated by two processes: 1) primary emissions from fossil fuel combustion and 2) mixing and dilution. Combustion of gasoline and diesel fuel is the primary source of CO and BCA, respectively, in Mexico City (Baumgardner et al., 2002). The onset of major vehicular movement is at 0600 LST and continues unabated throughout the day until approximately 2000 (PROAIRE, 1997). Figures 4a and 5b show that CO and BCM reach their peak at 0800 LST. Further increase is limited by the growth of the boundary layer after sunrise due to surface heating (Whiteman et al., 2000) that leads to turbulent

particle properties related to black carbon in Mexico City

D. Baumgardner et al.

Title Page

Abstract

Introduction

Conclusions

References

Tables

Figures

I◀

▶I

◀

▶

Back

Close

Full Screen / Esc

Printer-friendly Version

Interactive Discussion

mixing and dilution of gases and particles (Fast et al., 1998; Raga et al., 1999). The decrease observed in the CO and BCM after 0900 LST is a result of this process.

Although the maximum period of traffic is from 0600 to 1800 there is continuing vehicular activity 24 h a day, as is seen by the non-rush hour levels of CO and BCM of 0.5 ppm and 1000 ng m⁻³. The total particle concentration, represented by the CN measurements (Fig. 4b), follows a similar trend as CO and BCM. The CN rarely decreases below 10 000 cm⁻³ and only during the nighttime when day-to-day variability is at a minimum. Of all the environmental variables measured only the PPAH* decreases to values below its detection limit of 5 ng m⁻³. The PPAH* reaches a maximum at the same time as the CN, indicating that the majority of the particles containing PAH are quite small, i.e. those that dominate the CN concentration.

The average mass equivalent diameter (Fig. 5c) is inversely correlated with BCM, reaching a minimum when BCM is at its maximum. This is because the amount of BCM in individual particles is smaller during the initial increase of emissions in the early morning, but gradually increases during the day. The BCM_{fraction} (Fig. 5d) increases in the morning during the onset of traffic then decreases throughout the afternoon and evening. This suggests that when the average mass diameter is smaller, the particles have a thinner layer of non-light absorbing material than later in the day when they age and gases condense onto their surfaces. As seen in Figs. 5e and f, the relative fraction of particles that contain BC and the relative amount of total BCM increase at the same time as the BCM. The fraction of BCM containing particles increases by a factor of four from 10% to 40% from 0600 to 1000 am LST and the BCM increases by a factor of six, from 2% to 12% of the total mass.

The directly measured optical properties, B_{scat} and B_{abs}, are also linked to the total number concentration and BCM. Note that whereas B_{abs} reaches its maximum at the same time as CN (Fig. 4d), the B_{scat} reaches a peak two hours later. This illustrates the counterbalancing influences of particle growth versus mixing and dilution, i.e. light scattering is proportional to the number concentration and the optical cross section of particles. The number concentration of particles is decreasing by dilution but the

**particle properties
related to black
carbon in Mexico City**

D. Baumgardner et al.

Title Page

Abstract

Introduction

Conclusions

References

Tables

Figures

◀

▶

◀

▶

Back

Close

Full Screen / Esc

Printer-friendly Version

Interactive Discussion

average diameter is increasing through condensation and coalescence. The dilution process eventually overcomes the particle growth but delays the time of the maximum B_{scat} by approximately two hours.

5 The distribution by size of the number concentration with and without BCM, BCM concentration and $\text{BCM}_{\text{fraction}}$ are presented in Figs. 6a–d. The measurements were separated into four, six hour time periods and averaged over all days. The four curves in each panel are over the intervals 0000–0600, 0600–1200, 1200–1800 and 1800 to 2400. We observe that the size distributions of particles with no BCM have a different diurnal trend than those particles with BCM. The non-BCM particles increase in concentrations over all sizes between the first and second time period, maintaining their
10 peak at 250 nm. During the third time period the particles have grown and the peak of the distribution is at 350 nm where it remains into the evening, although at a lower concentration. The size distribution of the particles containing BCM (Fig. 6b) changes from a spectrum with a rather broad peak between 250 and 300 nm to a sharply peaked
15 spectrum in the second time period with maximum at 250 nm. In the following time periods the concentration is a factor of two lower and shifts slightly towards larger sizes. The BCM distribution (Fig. 6c) follows the same trend as the number concentration of BCA.

20 The $\text{BCM}_{\text{fraction}}$ as a function of size shows a distinct shift in the shape of the distributions between the morning and afternoon periods. In the morning, the smallest particles have the largest fraction of BCM in them, i.e. the thinnest coating of non-light absorbing material. In the afternoon, the peak has shifted from an average diameter of 220 nm to 300 nm.

7 Discussion

25 A more detailed examination of the diurnal trends provides further insight with respect to the underlying processes that control the evolution of aerosol properties in Mexico City. The increase in the mass equivalent diameter of BCM (Fig. 5c), after its minimum

**particle properties
related to black
carbon in Mexico City**

D. Baumgardner et al.

Title Page

Abstract

Introduction

Conclusions

References

Tables

Figures

◀◀

▶▶

◀

▶

Back

Close

Full Screen / Esc

Printer-friendly Version

Interactive Discussion

at 0800 am, is unlikely a result of coalescence, a process that will increase the size of BC once it has been produced from combustion, since this process requires many hours to cause a significant shift in the size distribution. A more reasonable explanation is that there are changes in the relative intensity of emission sources. Diesel combustion is the primary source of BCM in particles (Baumgardner et al., 2002); however, some BCM is produced from gasoline and the charcoal stoves that are used for cooking by street vendors. A shift from smaller, primary BCA to larger ones would suggest a shift from primarily auto traffic to a greater frequency of heavy vehicles like trucks and buses that use diesel. A higher emission rate of BCA would also increase the coalescence rate and lead to larger average mass diameters.

The rapid decrease of PPAH (Fig. 4c), within four hours of its maximum, suggests a removal process related to photochemical reactions; however, this behavior has been previously reported (Marr et. al., 2006) and it was concluded that the rapid decay in PPAH is a result of the coating of PAH-containing particles with secondary organics. This suppresses the response of the PAS 2000; hence, not only is the evolution of PPAH not linked to the dilution, as is the case with related primary emission variables, but it also urges caution in using this type of analyzer for analyzing urban PPAH.

The CN and B_{abs} reach their maxima simultaneously (Fig. 4) but CN decreases more rapidly than the B_{abs} . The matching peaks in B_{abs} and CN are expected since light absorption in Mexico City is primarily from high concentrations of BCA whose source is identical to the CN particles. The slower rate of decrease in B_{abs} (Fig. 4d) compared to that of the BCA concentration and mass (Fig. 5a and 5b) is somewhat puzzling at first inspection. The response of B_{abs} should be similar to CN and BMC, yet the afternoon trend follows that of B_{scat} (Fig. 4e). Recalling that B_{abs} was corrected for the interference of light scattering on the transmission through its filter and under the assumption that the B_{abs} values were biased by scattering from the filter, we adjusted the correction factor in (4) until the trend in B_{abs} more closely matched that of CN and BMC. The corrected values of B_{abs} matched the diurnal pattern of BMC when the correction factor was changed from 0.02 to 0.2. This is a significant difference from the

particle properties related to black carbon in Mexico City

D. Baumgardner et al.

Title Page

Abstract

Introduction

Conclusions

References

Tables

Figures

◀

▶

◀

▶

Back

Close

Full Screen / Esc

Printer-friendly Version

Interactive Discussion

correction factor suggested by Bond et al. (1999) and suggests several possibilities: 1) the aerosols that were used in the Bond et al. laboratory tests have significantly different optical properties than atmospheric aerosols in Mexico City, 2) the scattering coefficients measured by the nephelometer in the current study are too low by a factor of 10, or 3) a combination of both. One method to test the fidelity of the corrected B_{abs} is to compare values of the BC mass specific absorption, σ_{abs} , derived from the uncorrected and corrected B_{abs} . The σ_{abs} is defined as the absorption cross section per unit mass in units of $\text{m}^2 \text{g}^{-1}$ and is calculated by dividing B_{abs} by BCM. Figure 7 is a scatter plot that compares hourly averages of BCM with the corrected and uncorrected B_{abs} . The regression of BCM versus B_{abs} is constrained to pass through the origin under the assumption that the absence of BCM implies negligible B_{abs} . The slope of the regression is the value of σ_{abs} . With no correction, the derived σ_{abs} is $15.2 \text{ m}^2 \text{g}^{-1}$ whereas after the corrected σ_{abs} is $10.8 \text{ m}^2 \text{g}^{-1}$. Many studies have evaluated the relationship of σ_{abs} to the properties of BCM (e.g., Liousse et al., 1993; Petzold et al., 1997; Fuller et al., 1999) and have concluded that σ_{abs} can vary from $4 \text{ m}^2 \text{g}^{-1}$ to greater than $20 \text{ m}^2 \text{g}^{-1}$. Recent studies (Schuster et al., 2005; Barnard et al., 2005), using several independent methods to derive σ_{abs} , estimate a value between 8 and $10 \text{ m}^2 \text{g}^{-1}$ for Mexico City. Not only is the σ_{abs} derived from the corrected B_{abs} closer to the Schuster et al. and Barnard et al. studies, but the correction reduces the dispersion between BCM and B_{abs} .

A second method was used to assess the validity of B_{abs} corrections. As described previously, the optical properties of individual particles were estimated from the SP2 measurements using the derived diameter and $\text{BC}_{\text{fraction}}$ to calculate the relative fraction of BC and thickness of the coating material. Figures 8a–c compare the absorption and scattering coefficients, as well as the single scattering albedo, determined directly from the nephelometer and PSAP (dashed curves) and indirectly from the SP2 measurements (solid curves). These curves are the daily averages over the 14 days of sampling and show that the corrected B_{abs} from the PSAP (Fig. 8a) are in good agreement with the B_{abs} estimated from the SP2. This leads us to conclude that the addi-

particle properties related to black carbon in Mexico City

D. Baumgardner et al.

Title Page

Abstract

Introduction

Conclusions

References

Tables

Figures

◀

▶

◀

▶

Back

Close

Full Screen / Esc

Printer-friendly Version

Interactive Discussion

tional correction to B_{abs} is justifiable, although further investigation into the nature of the correction may be warranted to evaluate if this is unique to the aerosol properties in Mexico City. The directly measured and derived B_{scat} are in best agreement during periods when they are a maximum (Fig. 8b); however, the derived B_{scat} is much larger than the B_{scat} measured with the nephelometer outside of these periods. This discrepancy may be a result of the assumptions used for the refractive index of the non-light absorbing component of the particles. Another explanation is that the nephelometer has a heated inlet whose purpose is to remove water from aerosols prior to measurement. This reduces the size of the particles with a subsequent reduction in the scattering. This also explains the need for the larger correction factor to B_{abs} and implies that the original correction factor of 0.02 (Bond et al., 1999) would have been sufficient had the measured B_{scat} been determined from particles before they were dried. The SP2 and PSAP measured ambient aerosols with no heated inlet.

The contribution of the BCA to the total light extinction is shown in Fig. 8d where the percentage contribution was derived by calculating the extinction coefficient from the SP2 measurements with and without the particles that contained BCM. During those periods where 40% of the particles contained BC (Fig. 5e) and 12% of the total mass was BC, these particles contributed 20% to the total extinction. Figure 9 shows the size distributions of the extinction coefficient divided into four time periods. Here we see that the majority of the extinction is from particles larger than 300 nm, even though less approximately 60% of the BCM is in this size range. This highlights the radiative importance of BCM in the size range from 200 nm to 700 nm.

The CO concentrations are useful for diagnosing BCM when measurements of BCM are not available (Baumgardner et al., 2002). As shown in Fig. 10 the BCM in Mexico City is related to the CO by a factor of 1800, i.e. one ppm of CO is associated with 1800 ng m^{-3} of BC. When adjusted to standard atmospheric pressure and temperature, and converting ppm to mg m^{-3} , the BCM to CO relationship is one microgram of BCM for every milligram of CO. This is the same relationship that was previously derived by Baumgardner et al. (2002). According to the 1999 inventory of emissions (PROAIRE,

particle properties related to black carbon in Mexico City

D. Baumgardner et al.

Title Page

Abstract

Introduction

Conclusions

References

Tables

Figures

◀

▶

◀

▶

Back

Close

Full Screen / Esc

Printer-friendly Version

Interactive Discussion

1999), Mexico City produces approximately 1.2×10^6 metric tons of CO yearly. Using the BCM to CO relationship we predict 1200 metric tons of BCM emitted yearly in Mexico City. This is similar to the value of 1700 ± 200 metric tons predicted by Jiang et al. (2006) who used measurements of B_{abs} to estimate BCM.

5 **8 Summary**

Measurements of the BCM in individual, internally mixed aerosol particles were made with the SP2 in April of 2003 and 2005. The measurement technique uses light scattering and incandescence to derive the optical and BC mass equivalent diameters from which the fraction of BCM in each particle can be estimated. The scattering and absorption coefficients were also derived from the SP2 measurements and compared with these optical properties that were directly measured with a nephelometer and soot photometer. The characteristics of the BCA were compared with the CO, CN, and PPAH concentrations, those environmental variables that are also linked to primary emissions.

15 The BCM, CO and CN have matching diurnal trends that are associated with traffic patterns and boundary layer growth. The BCM reaches a daily average maximum of $5000 \pm 1000 \text{ ng m}^{-3}$ but has a nighttime, background level of 1000 ng m^{-3} . The number of particles containing BCM ranges between 10% to 40% of all particles between 200 nm and 700 nm and the BCM is between 4% and 12% of the total mass in this size range. The BC particles have the thinnest coating of non-light absorbing material during periods of maximum BCM.

20 The absorption coefficient derived from the SP2 was in close agreement with direct measurements made with the PSAP. Approximately 60% of the BCM and 80% of the extinction coefficient are in particles with diameters larger than 200 nm. BCM is emitted at a rate of 1200 metric tons per year in Mexico City, based upon the SP2 measurements and correlations between BCM and CO.

The properties of the BCA that have been provided by this study improve our under-

**particle properties
related to black
carbon in Mexico City**

D. Baumgardner et al.

Title Page	
Abstract	Introduction
Conclusions	References
Tables	Figures
◀	▶
◀	▶
Back	Close
Full Screen / Esc	
Printer-friendly Version	
Interactive Discussion	

standing of how aerosols evolve in urban areas. These properties should be taken into account in applications that require accurate depictions of atmospheric aerosols.

Acknowledgements. The SP2 was developed under Navy contract #N00014-01-C-0335. We would like to thank A. Retama of the Mexico City environmental service for providing the CO data from the RAMA network.

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particle properties
related to black
carbon in Mexico City

D. Baumgardner et al.

Title Page

Abstract

Introduction

Conclusions

References

Tables

Figures

◀

▶

◀

▶

Back

Close

Full Screen / Esc

Printer-friendly Version

Interactive Discussion

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particle properties related to black carbon in Mexico City

D. Baumgardner et al.

Title Page

Abstract

Introduction

Conclusions

References

Tables

Figures

◀

▶

◀

▶

Back

Close

Full Screen / Esc

Printer-friendly Version

Interactive Discussion

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particle properties related to black carbon in Mexico City

D. Baumgardner et al.

Title Page

Abstract

Introduction

Conclusions

References

Tables

Figures

◀

▶

◀

▶

Back

Close

Full Screen / Esc

Printer-friendly Version

Interactive Discussion

- Slowik, J. G., Cross, E., Han, J., Davidovits, P., Onasch, T. B., Jayne, J. T., Williams, L. R., Canagaratna, M. R., Worsnop, D. R., Chakrabarty, R. K., Arnott, W. P., Schwarz, J. P., Gao, R. S., Fahey, D. W., and Kok, G. L.: Intercomparison of instruments measuring black carbon content and optical properties of soot particles, *Aerosol. Sci. Technol.*, accepted, 2007.
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**particle properties
related to black
carbon in Mexico City**

D. Baumgardner et al.

Title Page

Abstract

Introduction

Conclusions

References

Tables

Figures

◀

▶

◀

▶

Back

Close

Full Screen / Esc

Printer-friendly Version

Interactive Discussion

**particle properties
related to black
carbon in Mexico City**

D. Baumgardner et al.

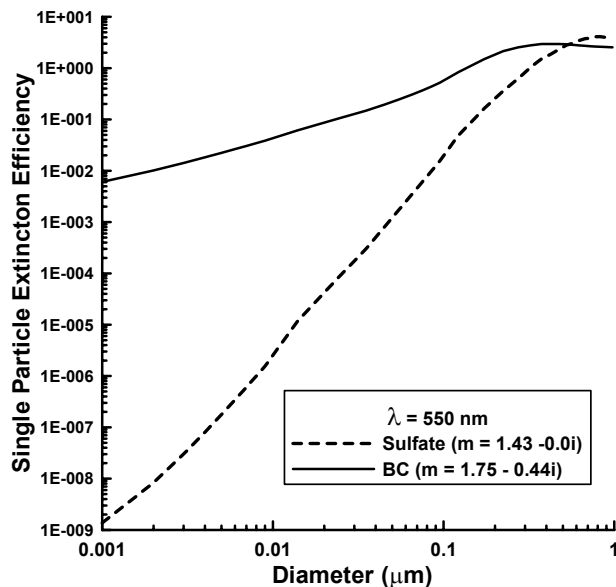


Fig. 1. The extinction efficiency of BC (solid) is compared with that of ammonium sulfate (dashed).

[Title Page](#)[Abstract](#)[Introduction](#)[Conclusions](#)[References](#)[Tables](#)[Figures](#)[I◀](#)[▶I](#)[◀](#)[▶](#)[Back](#)[Close](#)[Full Screen / Esc](#)[Printer-friendly Version](#)[Interactive Discussion](#)

**particle properties
related to black
carbon in Mexico City**

D. Baumgardner et al.

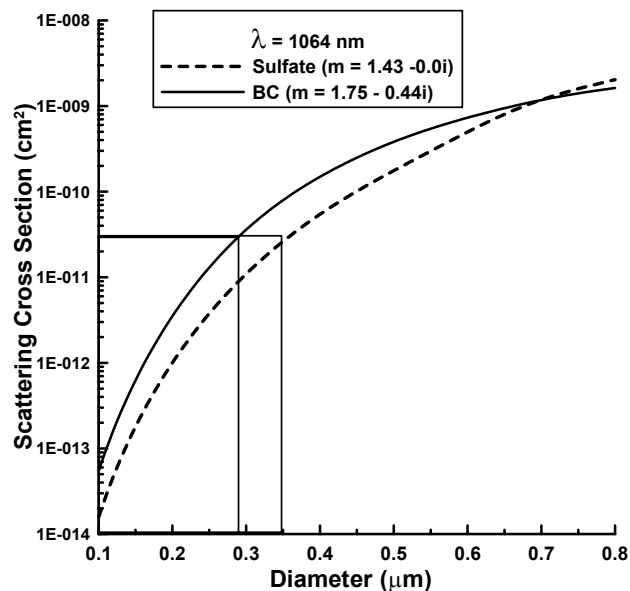


Fig. 2. Scattering cross sections for sulfate (dashed) and BC (solid) particles as a function of particle size for the collection angles (30°–60° and 120°–150°) and wavelength of the SP2 laser.

[Title Page](#)[Abstract](#)[Introduction](#)[Conclusions](#)[References](#)[Tables](#)[Figures](#)[◀](#)[▶](#)[◀](#)[▶](#)[Back](#)[Close](#)[Full Screen / Esc](#)[Printer-friendly Version](#)[Interactive Discussion](#)

**particle properties
related to black
carbon in Mexico City**

D. Baumgardner et al.

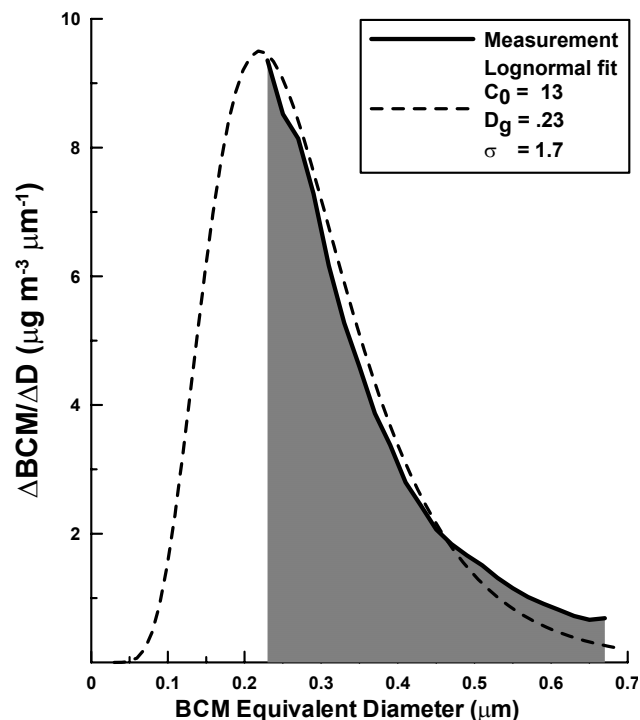


Fig. 3. The BCM in particles smaller than the size threshold of the SP2 (0.20) is estimated by fitting a lognormal distribution to the measurements (dashed curve). As compared to the total BCM estimated from the lognormal distribution, the SP2 measures approximately 60% of the ambient BCM as shown by the shaded area under the measured size distribution (solid curve).

[Title Page](#)[Abstract](#)[Introduction](#)[Conclusions](#)[References](#)[Tables](#)[Figures](#)[◀](#)[▶](#)[◀](#)[▶](#)[Back](#)[Close](#)[Full Screen / Esc](#)[Printer-friendly Version](#)[Interactive Discussion](#)

**particle properties
related to black
carbon in Mexico City**

D. Baumgardner et al.

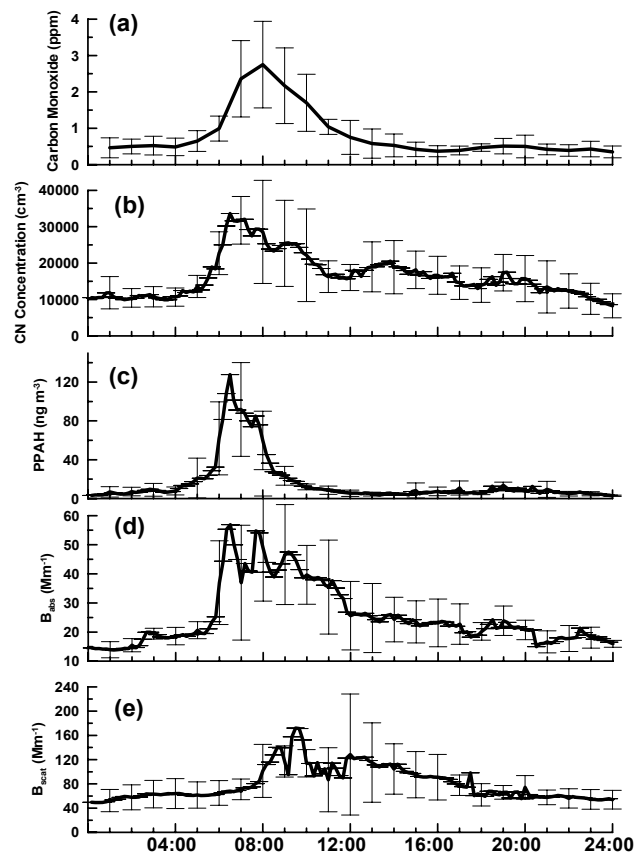


Fig. 4. Typical diurnal evolution derived from averaging all sampling days, for **(a)** CO, **(b)** CN, **(c)** PPAH*, **(d)** B_{abs} and **(e)** B_{scat} . The vertical bars indicate one standard deviation and represent an estimate of the day-to-day variability.

[Title Page](#)[Abstract](#)[Introduction](#)[Conclusions](#)[References](#)[Tables](#)[Figures](#)[◀](#)[▶](#)[◀](#)[▶](#)[Back](#)[Close](#)[Full Screen / Esc](#)[Printer-friendly Version](#)[Interactive Discussion](#)

**particle properties
related to black
carbon in Mexico City**

D. Baumgardner et al.

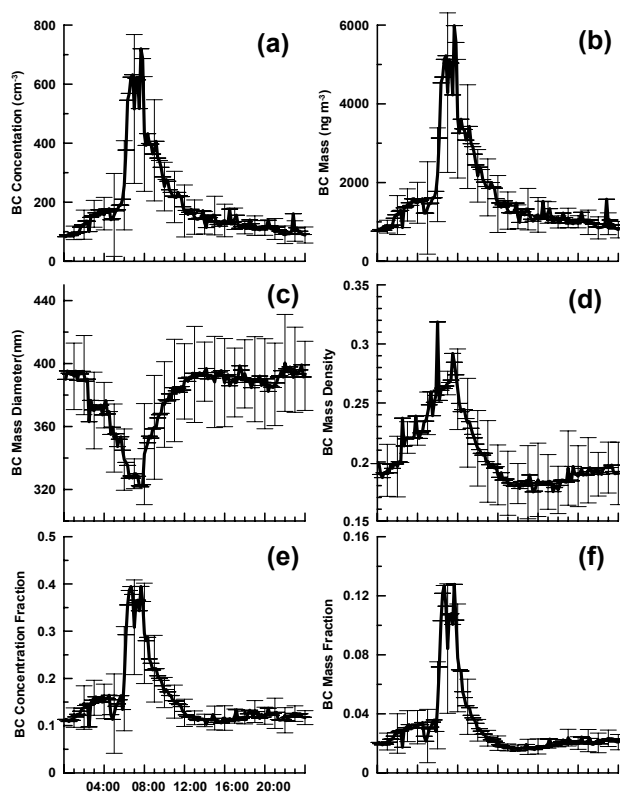


Fig. 5. As in Fig. 4 but for (a) BC concentration, (b) BC mass, (c) $D_{\text{mass-eqv}}$, (d) BC mass density, (e) BC fraction and (f) BC mass fraction. The vertical bars indicate one standard deviation and represent an estimate of the day-to-day variability.

Title Page

Abstract

Introduction

Conclusions

References

Tables

Figures

◀

▶

◀

▶

Back

Close

Full Screen / Esc

Printer-friendly Version

Interactive Discussion

**particle properties
related to black
carbon in Mexico City**

D. Baumgardner et al.

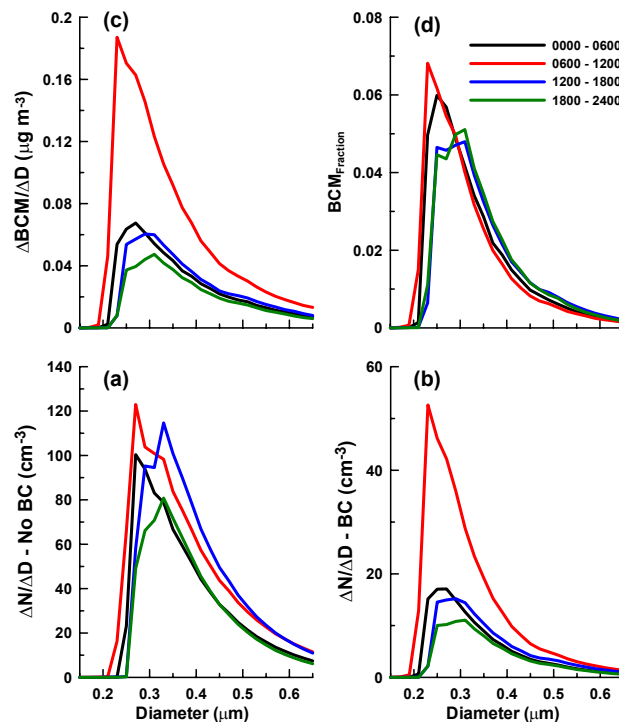


Fig. 6. These size distributions are daily averages every six hours for the 14 days of sampling and show the number concentrations of particles with no BC (a), with BC (b), the BC mass concentration (c) and the BCM fraction (d).

[Title Page](#)[Abstract](#)[Introduction](#)[Conclusions](#)[References](#)[Tables](#)[Figures](#)[◀](#)[▶](#)[◀](#)[▶](#)[Back](#)[Close](#)[Full Screen / Esc](#)[Printer-friendly Version](#)[Interactive Discussion](#)

**particle properties
related to black
carbon in Mexico City**

D. Baumgardner et al.

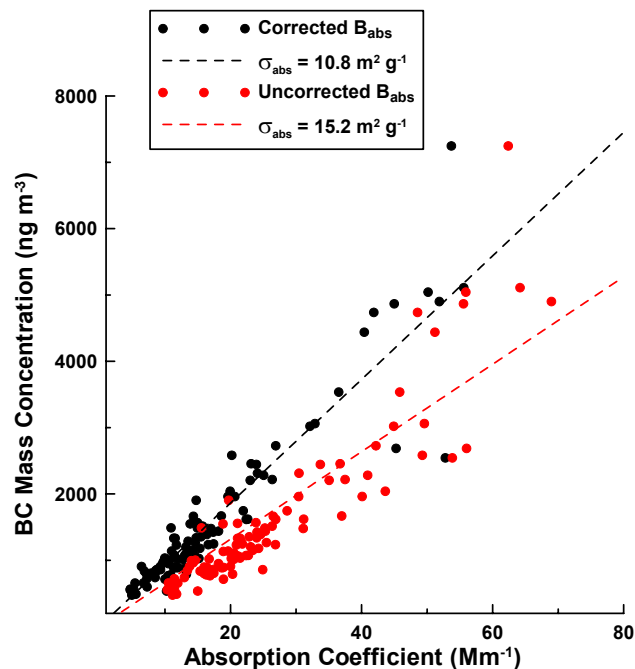


Fig. 7. The BC mass concentration, compared with the absorption coefficient, provides an estimate of the BC mass specific absorption coefficient, σ_{abs} . Here the two values of σ_{abs} are derived from the uncorrected (red) and corrected (black) directly measured values of B_{abs} .

Title Page

Abstract

Introduction

Conclusions

References

Tables

Figures

◀

▶

◀

▶

Back

Close

Full Screen / Esc

Printer-friendly Version

Interactive Discussion

particle properties
related to black
carbon in Mexico City

D. Baumgardner et al.

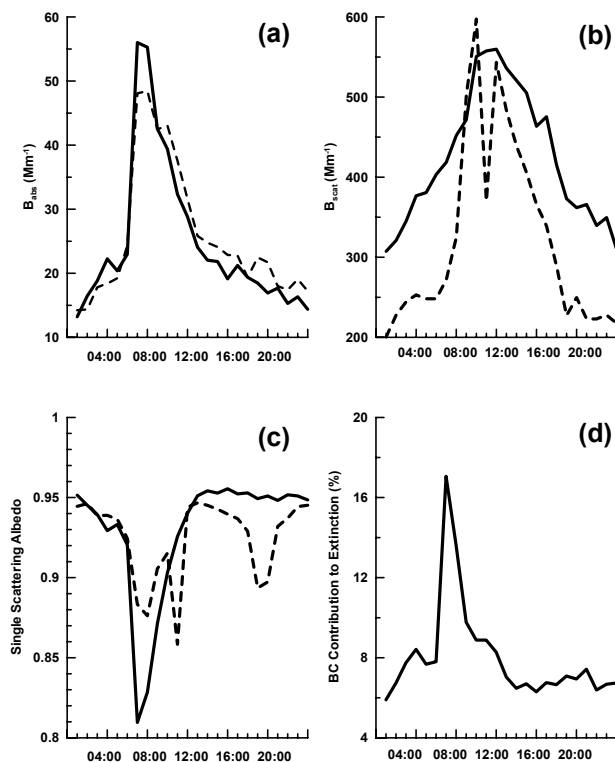


Fig. 8. The evolution of the particle optical properties, derived from averaging all sampling days, is compared here for B_{abs} (a), B_{scat} (b), and single scattering albedo (c), derived indirectly from the SP2 measurements (solid) and directly from the nephelometer and PSAP (dashed). The contribution of the BC aerosol to the total extinction is shown in panel (d).

[Title Page](#)[Abstract](#)[Introduction](#)[Conclusions](#)[References](#)[Tables](#)[Figures](#)[◀](#)[▶](#)[◀](#)[▶](#)[Back](#)[Close](#)[Full Screen / Esc](#)[Printer-friendly Version](#)[Interactive Discussion](#)

**particle properties
related to black
carbon in Mexico City**

D. Baumgardner et al.

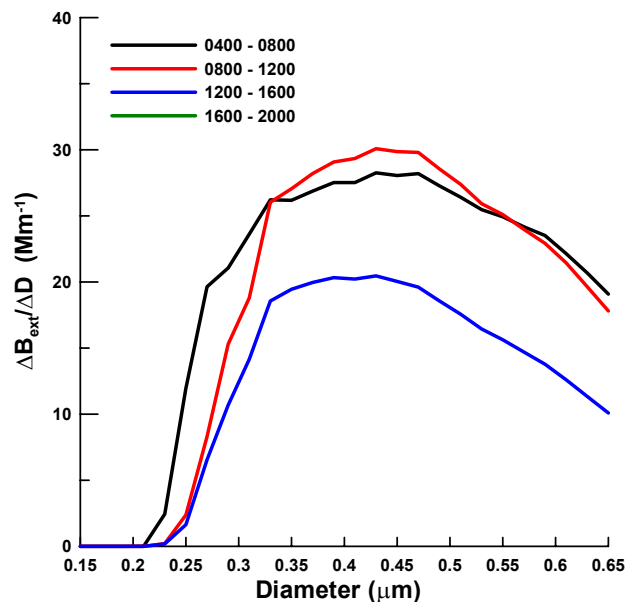


Fig. 9. These size distributions of the extinction coefficient derived from the SP2 measurements are daily averages every six hours for the 14 days of sampling.

[Title Page](#)[Abstract](#)[Introduction](#)[Conclusions](#)[References](#)[Tables](#)[Figures](#)[◀](#)[▶](#)[◀](#)[▶](#)[Back](#)[Close](#)[Full Screen / Esc](#)[Printer-friendly Version](#)[Interactive Discussion](#)

**particle properties
related to black
carbon in Mexico City**

D. Baumgardner et al.

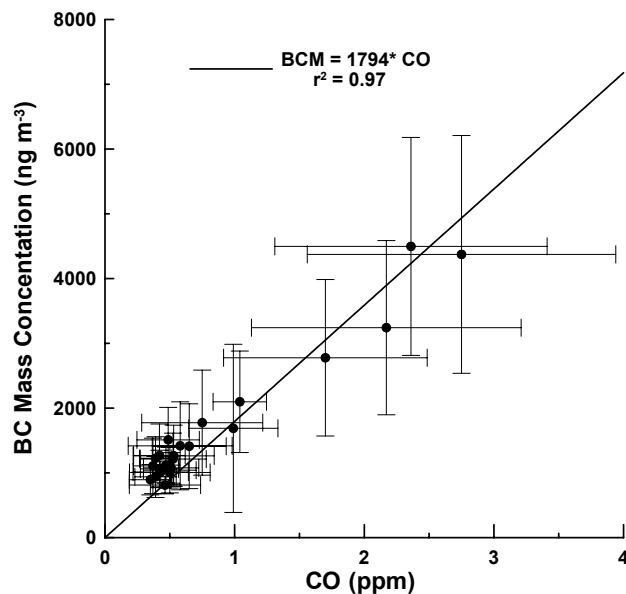


Fig. 10. Relationship between BC mass and CO derived from present study, and similar to that obtained in (Baumgardner et al., 2002).

[Title Page](#)[Abstract](#)[Introduction](#)[Conclusions](#)[References](#)[Tables](#)[Figures](#)[◀](#)[▶](#)[◀](#)[▶](#)[Back](#)[Close](#)[Full Screen / Esc](#)[Printer-friendly Version](#)[Interactive Discussion](#)